

Research Article

# Vertical distribution of the salt marsh invader *Spartina alterniflora* and native halophytes on the west coast of Korea in relation to tidal regimes

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#### Abstract

Smooth cordgrass (Spartina alterniflora Loisel.), an aggressive non-native species worldwide, colonized tidal flats on the west coast of Korea in two regions differing in tidal amplitude between 1990–2004. By the time of our study in 2015, expansion had occurred both clonally and through formation of new patches, providing an opportunity to determine intertidal range, which is a key component of understanding the threat posed by S. alterniflora through competition with native halophytes or transformation of unstructured mudflat. At Ganghwa (5.69 m tidal range), S. alterniflora ranged from 3.52 to 1.34 m above Mean Sea Level (MSL). At Jindo (2.02 m tidal range), S. alterniflora ranged from 1.57 to -0.18 m relative to MSL. Thus, a wider absolute intertidal range was occupied by S. alterniflora at the megatidal vs mesotidal region, but the lower limit of S. alterniflora did not extend below MSL under megatidal conditions, a pattern that now appears to emerge consistently in both the native and introduced range. In both study regions, S. alterniflora occurred at the same elevations as other salt marsh plants, occupying an upper zone with Phragmites australis (non-native) and middle zone with several native species including Suaeda japonica. S. alterniflora occurred below native marsh vegetation at all sites, which would result in transformation of the extensive mudflats along the Korean coast.

**Key words:** saltmarsh plants, invasive species, megatidal flat, Yellow Sea, Korean coastal wetland, Ganghwa, Jindo

# Introduction

Salt marshes are characterized by striking zonation of vascular plants correlated with marsh elevation (Chapman 1974; Gray 1992). Lower limits are typically defined by physical stress of salinity or waterlogging, and few plant species have adaptations that enable survival below the high tide mark. One such species is

smooth cordgrass (*Spartina alterniflora*, hereafter *Spartina*), which accordingly has had transformative effects after introduction to tidal flats lacking vegetation at that zone (Strong and Ayres 2013; Ruesink 2018). *Spartina*'s intertidal distribution in its native range tends to increase with tidal amplitude, with an additional latitudinal factor appearing to restrict its upper limit (McKee and Patrick 1988). The recent invasion of *Spartina* along the west coast of Korea allows evaluation of whether these patterns of intertidal distribution persist in novel meso- and megatidal regions and overlap with native marsh.

*Spartina*, native to the Atlantic and Gulf coasts of America (Simenstad and Thom 1995; Blum et al. 2007), exclusively dominates low marsh habitats, followed by dense, monospecific stands of *Spartina patens* at slightly higher elevations, before reaching a suite of species in the high marsh zone (Bertness 1991). *Spartina* generally grows between Mean High Water (MHW) and Mean Low Water (MLW). However, no consistent elevation exists for zonation relative to a tidal datum due to biotic interactions, tidal inundation, and local or regional differences in other factors, such as salinity and physical disturbance (McKee and Patrick 1988).

*Spartina* has invaded in the western United States, Australia, New Zealand, Japan, and China, following either intentional or unintentional introduction (An et al. 2007; Maebara et al. 2020). The growth form includes colonization by seeds or fragments and subsequent clonal expansion into dense patches that ultimately fuse into meadows. Stems can accumulate sediment and/or build up rhizome mats below ground that effectively raise the level of the tidal flat. *Spartina* has a wide ecological niche, occupying sediments ranging from silt to cobble, a large latitudinal range (Mexico to Nova Scotia where native; Bruno and Kennedy 2000; Pennings and Bertness 2001), and salinity from brackish to saltwater (Nestler 1977; Field observation from Korean coast). Such high environmental tolerance may enable still wider distribution where introduced, as the realized niche is less restricted, which in turn can lead to competition with native vegetation. Additionally, genetic changes appear to have occurred in *Spartina* invading China and Japan that may enhance performance in the invaded range (Liu et al. 2020).

Two regions along the west coast of Korea currently have *Spartina*. The introduction of *Spartina* was reported in 2015 (Jung et al. 2015; Kim et al. 2015). These were reported as *S. anglica* in Ganghwa and *S. alterniflora* in Jindo respectively, later revealed as *S. alterniflora* through total genomic DNA studies from these two study sites (Hong JS, unpublished data). At Ganghwa, *Spartina* is estimated to have arrived at Buno in 1990 and Dongmak in 2004. At Jindo, *Spartina* is estimated to have arrived at the local site of Namdong-ri in 2004. As of 2015, *Spartina* occupied 2,245 m<sup>2</sup> in Buno, 8,738 m<sup>2</sup> at Dongmak, and 6,263 m<sup>2</sup> at Namdong-ri (unpubl. Data by authors). By 2020, *Spartina* had expanded to over 40,000 m<sup>2</sup> of previously unvegetated tidal flat in at least eight different sites along the west coast of Korea (Jung HI, Korea Marine Environment Management Corporation, personal communication), despite substantial investment (3.5B Korean Won = US \$3 million) in mechanical control efforts of plowing and mowing.

The goal of this study was to evaluate the extent of transformation of tidal mudflats on the west coast of Korea likely to occur from *Spartina*, and the extent of overlap with native marsh species. To accomplish this goal, we accurately measured the vertical distribution of *Spartina* and native halophytes at sites with different tidal regimes. We expected that the intertidal distribution of invasive *Spartina* would at least match that of the native range and potentially be wider.



# Materials and methods

#### Study sites

Two regions on the west coast of South Korea about 400 km apart have been invaded by Spartina marshes. Ganghwa is located at the estuaries of the Han River near the Seoul metropolitan area. The southern coast of Ganghwa Island has a wide open-coast megatidal muddy flat with a total area of about 105 km<sup>2</sup> (approximately  $17.5 \times 6$  km, Fig. 1). A large amount of sediment is supplied from the land and transported by tide and currents with the mixing of Yellow Sea seawater and freshwater from the Han River (Lee et al. 2011). The mean tidal range is about 6 m (KHOA 2017). Two sites within the Ganghwa region were known to contain Spartina in 2015. At the first site, Dongmak (37°35.62'N, 126°26.74'E), the upper zone consisted of small patches of reed marsh (*Phragmites australis*) and scattered other halophytes (Triglochin maritimum and Schoneoplectus triqueter). Further down, Suaeda japonica occurred in a dense band of 100 m width. In 2015, eleven years after its introduction in 2004, Spartina occurred along a 1.5 km section of coastline, with small patches, many of which had expanded and fused. At the second site, Buno (37°35.51'N, 126°27.76'E), where Spartina likely arrived in 1990, a large Spartina patch was located adjacent to the dock in a relatively enclosed area, and small patches were distributed along the coastline to the north. No native vegetation was present at this site. The second region of Korea invaded by Spartina is in the southwestern part of Jindo Island on the southernmost part of the west coast of Korea, where the tidal range is about 2 m (KHOA 2017). The study site, Namdong-ri (34°21.86'N, 126° 9.72'E), is a small tidal flat of about 0.029 km<sup>2</sup> in an enclosed bay, with an entrance just 50 m wide. The upper part of the bay is connected to a rivulet from the land. Spartina vegetation, which first appeared in 2004, was by 2015 composed of rather cohesive and atypical large patches. The largest Spartina patch was about 100 m wide. The upper zone was dominated by halophytes such as P. australis, Zoysia sinica, and Suaeda maritima.

# Field survey

Field surveys were conducted at Jindo and Ganghwa tidal flats in June and August 2015, respectively. Native halophytes and Spartina inhabited these tidal flats, including in front of the levee in both regions. To measure vertical distribution, we included all patches along each transect. Transects were established from the upper part of the tidal flat past the lower extent of any vegetation. They were 15 m wide and perpendicular to the shore, extending 300 m at Dongmak, 50 m at Buno, and 100 m at Namdong-ri. Five transects were established at Dongmak and Namdong-ri, and two at Buno. Intertidal elevation was determined at stations every 1-10 m, with closer intervals where slope condition and marsh plant presence changed (sextant-based positioning). Distance between stations was recorded, and at each station, tidal elevation was determined (Total Station, PANTAX R-300) in relation to nearby spatial reference points (National Geographic Information Institute in Korea - NGII and Korea Hydrographic and Oceanographic Agency -KHOA). We used the harmonic constants obtained from KHOA at each site to extract elevation for the following tidal datums: AHHW, MHWS, MHW, MHWN, MTR (MHW–MLW) (Table 1). We used the term Half Tide Level (HTL) to refer to the plane midway between MHW and MLW, which is similar to Mean Sea Level (McKee and Patrick 1988).



| Tidal datum                     | Tidal datum abbreviation | Ganghwa (Dongmak, Buno) | Jindo (Namdong-ri) |
|---------------------------------|--------------------------|-------------------------|--------------------|
| Approximately Higher High Water | AHHW                     | 4.67                    | 1.95               |
| Mean High Water Spring          | MHWS                     | 3.99                    | 1.35               |
| Mean High Water                 | MHW                      | 2.85                    | 1.01               |
| Mean High Water Neap            | MHWN                     | 1.73                    | 0.66               |
| Mean Tidal Range (MHW–MLW)      | MTR                      | 5.69                    | 2.02               |

Table 1. Mean tidal range and elevations of tidal datums in each study site. Values represent the elevation above Mean Sea Level (m).

# Vegetation mapping and image analysis

Area of vegetation was determined from photographs at an altitude of 50 m using a drone (DJI Phantom 3) equipped with a digital camera. Prior to photography, reference points for measuring GSD (Ground Sample Distance) were arbitrarily set and their geopositions recorded. Multiple photos from the drone were combined to create a vegetation map for each 15 m-wide transect using Agisoft PhotoScan pro, an image editing software. Elevations measured in the field were adjusted to local elevation (meters above local mean sea level) and incorporated as a data layer along each transect. Based on the map, we described vegetation type, percent coverage, and distribution area per 0.1 m interval of elevation.

# Comparison with global data on intertidal distribution of Spartina alterniflora

We found reports of *Spartina* upper and lower limits at 28 sites within the native range that differed in tidal regime, as well as seven sites outside the native range. We compiled this information to assess whether the intertidal distribution in Korea showed upper and lower limits consistent with the meso- and mega-tidal regimes along the coast.

# Results

# Tidal condition, profile, and vertical distribution of salt marshes

Mean tidal range (MTR) at Ganghwa was 5.69 m, more than twice that of 2.02 m in Jindo (Table 1). The slope of the tidal flat was shallowest at Dongmak (Ganghwa) (Fig. 2A, B). Namdong-ri site in Jindo was relatively steep, with high bathymetric variation in the upper tidal flat and lower parts near the tidal creek (Fig. 2C). There were tidal creeks at the end of the marsh, coinciding with the boundary of the *Spartina* vegetation (Figs 1, 2C).

Marshes including *Spartina* vegetation of Dongmak were vertically distributed from 3.63 m to 1.49 m ( $\Delta$  Elevation = 2.14 m) above MSL. Those of Buno were vertically distributed from 2.70 m to 1.34 m ( $\Delta$  Elevation = 1.36 m), with the upper limit restricted due to the artificial wall. Those of Namdong-ri were distributed from 1.69 m to -0.18 m ( $\Delta$  Elevation = 1.87 m) (Fig. 2). Comparing tidal datums presented in Table 1 with Fig. 2, marshes at Ganghwa were distributed from below MHWS (Dongmak) or MHW (Buno) to below MHWN. However, in Namdong-ri of Jindo, marshes were distributed from below AHHW to below MSL, meaning that the marsh of Namdong-ri experienced a wider variety of tidal inundation duration than at Ganghwa.



Vertical distribution of Korean saltmarsh plants including Spartina



**Figure 1.** The study areas, Ganghwa and Jindo tidal flats located in the west coast of Korea and photographs showing the survey transects established in each sampling site in Ganghwa (Dongmak, D1–5; Buno, B1, 2) and Jindo (Namdong-ri, N1–5).

# Distribution of Spartina marsh by drone image analysis

Dongmak and Buno, Ganghwa

The areal cover of *Spartina* vegetation on transects of Dongmak was 1,918 m<sup>2</sup>, which was 14% of total area surveyed (Fig. 3A). Based on image analysis of the vegetation map, the coverage of *Spartina* patches here was the highest in transect D1 (28%) where the largest patch of *Spartina* occurred. The coverage of *Spartina* patches was the lowest in transect D4 (3%). The coverage increased rapidly below MHW, peaking at 2.5 m above MSL. It gradually decreased to MHWN. The mean elevation of *Spartina* marsh was 2.43 m above MSL. More than 90% of surveyed *Spartina* marshes were distributed between 2.9 m and 1.8 m. The highest patch appeared at 3.52 m while the lowest patch appeared at 1.49 m





Figure 2. The transect profiles of the study site. Solid trapezoid marks indicate an artificial levee (A Dongmak; B Buno; C Namdong-ri).

above MSL (Fig. 3A, Table 2). Therefore, the vertical distributional range of *Spartina* patches in Dongmak was 2.03 m. The highest patch was on transect D5. Its size was small. Transect D5 included a narrow channel about 3 m wide. The channel was connected to a small stream of freshwater input (Fig. 1). The patch on D5 transect was close to MHWS. In Buno, only the *Spartina* vegetation existed. The areal cover of the marsh on transects was 634 m<sup>2</sup>, accounting for 45% of the survey area (Fig. 3B). The mean elevation, main distributional range, and lower limit of the distribution were similar to those of the nearby Dongmak. However, the upper limit of the *Spartina* marsh was 2.70 m above MSL. Its distribution range was 1.36 m, which was narrower than that in Dongmak. However, the distribution was limited in the uppermost part due to the artificial levee (Figs 1, 2).





**Figure 3.** Vertical distribution of *Spartina* vegetation, native marsh and bare mudflat on the survey transects based on cover in aerial images. Total area estimates include the 12 belt transects surveyed. MSL: Mean Sea Level. See Table 1 for other tidal datums.

Namdong-ri, Jindo

The areal cover of *Spartina* vegetation on transects of Namdong-ri was 2,104 m<sup>2</sup>, accounting for 48% of total area surveyed (Fig. 3C). The mean elevation of the *Spartina* vegetation was 0.53 m (below MHWN) (Table 2), which was lower than the mean elevation of the *Spartina* vegetation in Ganghwa above MHWN. About 90% of the *Spartina* vegetation was distributed at 0.7 m to 0.2 m, which was MHWN to near MSL (Fig. 3C). The coverage was the highest at 0.7 m but similar (65–83%) within the main distribution range. It decreased rapidly near the tidal creek at 0.2 m above MSL. This abrupt decline contrasted with Ganghwa, where *Spartina* gradually decreased at lower elevations (Fig. 3). The highest patch appeared at 1.57 m while the lowest patch appeared at -0.18 m relative to MSL (Fig. 3C, Table 2). Therefore, the vertical distributional range of *Spartina* was 1.74 m.

# Native vegetation

Native vegetation was distributed from below MHWS to near MHWN in Dongmak and from below AHHW to MHWN in Namdong-ri (Fig. 4). Common reed, *P. australis*, was the highest vegetation in both regions. It was distributed below MHWS in Dongmak and below AHHW in Namdong-ri. Below the reed marsh, *T. maritimum*, *S. triqueter*, and *S. japonica* were distributed in Dongmak. *Z. sinica* and *S. maritima* were distributed in Jindo in order of elevation from high to low. In both regions, *Spartina* patches occurred throughout the elevational range of *P. australis*. In addition, *Spartina* occurred below all other vegetation in both regions. Therefore, vertical distribution ranges of *Spartina* were much larger than those of all native salt marshes surveyed. In terms of the *Spartina* invasion within the transect, *Spartina* marsh occupied 47% of the total area of marsh plants in

| Table 2. Sur | nmary of the v | vertical distributi | on of <i>Spartina</i> r | narsh in study site | s. Values repres | ent the elevation a | above Mean | Sea Level | (m). |
|--------------|----------------|---------------------|-------------------------|---------------------|------------------|---------------------|------------|-----------|------|
|--------------|----------------|---------------------|-------------------------|---------------------|------------------|---------------------|------------|-----------|------|

| Locality       | Gai       | Jindo     |            |
|----------------|-----------|-----------|------------|
| Site           | Dongmak   | Buno      | Namdong-ri |
| Upper limit    | 3.52      | 2.70      | 1.57       |
| Mean elevation | 2.43±0.31 | 2.30±0.31 | 0.53±0.26  |
| Lower limit    | 1.49      | 1.34      | -0.18      |
| Growth Range   | 2.03      | 1.36      | 1.74       |





Dongmak and 86% in Namdong-ri (Fig. 3). Invasion of *Spartina* into the zone of *P. australis* was clearly observed in D2 and D5 of Dongmak and N3 of Namdong-ri. In transect D2 of Dongmak, small patches of *Spartina* were attached to the edge of the *P. australis* marsh (Fig. 5). However, in D5, patches of *Spartina* and *P. australis* of similar sizes were attached. Between AHHW and MHW of N3 in Namdong-ri, *Spartina* patches existed inside *P. australis* patch as well as *S. maritima* (Fig. 5). Native vegetation was not distributed below MHWN, the main distribution range of the *Spartina* in Namdong-ri (Fig. 4). Therefore, *Spartina* of Namdong-ri existed alone in their main distribution range. On the other hand, the main distributional range of *Spartina* at Dongmak, MHW and MHWN, overlapped that of *S. japonica* (Fig. 4), a halophyte at the lower edge of native marsh. For that reason, most patches of *Spartina* in Dongmak were found in the *S. japonica* marsh.



#### A. Transect D2 Dongmak





# Comparison with global data on intertidal distribution of Spartina alterniflora

The intertidal elevations occupied by *Spartina* in 2015 at three sites in Korea were consistent with reports elsewhere in its native and non-native range, once accounting for tidal regime (Fig. 6). In much of the native range of *Spartina*, vertical distribution increases directly with tidal amplitude. However, above a tidal range of 3 m, the lower limit of *Spartina* rises approximately in parallel with the upper limit (Fig. 7). Meso-tidal and mega-tidal Korean sites thus showed primarily a shift upwards in *Spartina* as tidal range increased.





**Figure 6.** The elevational range of growth of *Spartina alterniflora* at locations along the east coast of America, San Francisco, east coast of China, east coast of Japan and the west coast of Korea, modified from McKee and Patrick (1988). 1: Kimura et al. 2016, 2: Callaway and Josselyn 1992, 3: present study, 4: Zhang et al. 2004, 5: Byers and Chmura 2007. HTL: Half Tide Level

# Discussion

# Vertical distribution range of Spartina in the study area

*Spartina*'s main distributional range and elevation can vary regionally. In their native regions, *Spartina* grows principally in between MHW and MLW (McKee and Patrick 1988; Bertness 1991), except shifted higher under macrotidal conditions in the Bay of Fundy (Byers and Chmura 2007). In invaded sites in Willapa Bay, US west coast, distribution was reported from MHHW to about 1 m above MLLW (Sayce 1988). In China, *Spartina* can grow between MHW and MSL on Jiangsu tidal flat of Yangtze River Estuary, where *Spartina* was planted in 1979 to promote conversion of tidal flats into dry land (Zhang et al. 2004; An et al. 2007; Meng et al. 2020). Our study showed that the *Spartina* marshes were mainly distributed between MHW and MSL. Thus, *Spartina* is located for the moment similar to vertical ranges where native and where introduced in the Yangtze Estuary.

*Spartina alterniflora* is well known as an estuarine species. It has a wide variety of temperature and salinity ranges, occupying substrates from cobble to sand and mud flat (Nestler 1977; Bruno and Kennedy 2000). These wide ecological niches may allow *Spartina* to occur beyond their general distributional ranges. In fact, we observed that marginal ranges of *Spartina* in both regions were much wider than

its main distributional ranges. Its patches were found even in the freshwater stream on the very upper intertidal area.

McKee and Patrick (1988) compiled vertical distributions of Spartina in the Atlantic and Gulf coast of America (Fig. 6). Its distribution in Jindo was measured with an upper limit between AHHW and MHWS, and lower limit below MSL. In the present study, the upper limit of the *Spartina*'s natural distribution in Jindo could not be measured due to the artificial levee. Nonetheless, compared to data from McKee and Patrick (1988), the distribution at Jindo appears similar in both Spartina distribution and tidal amplitudes to Kingsley Creek, Florida (Fig. 6). In Ganghwa, however, the upper limit of the distribution was between MHWS and MHW and the lower limit was close to MHWN, which was 1.34 m above MSL, and the regional comparisons for vertical distribution of Spartina showed a distribution rather at the upper part of the tidal elevation (Fig. 6). According to Fig. 6, our results look similar to Dipper Harbor and Saints Rest in the Bay of Fundy, Canada in native regions, and Badou and Wangzhu in East coast of China among invaded regions (Zhang et al. 2004; Byers and Chmura 2007), both of which have large tidal amplitude. The upper limit of the distribution was close to or above MHW, which is a typical upper limit across studies regardless of tidal amplitude. However, under meso- and mega-tidal regimes, the lower limit tends to shift well above MTL (Fig. 6). This pattern of vertical shift is different from what was previously reported for the native range on the east coast of United States, where elevational range increased in response to greater tidal amplitudes as Spartina extended both upper and lower boundaries (McKee and Patrick 1988) (Fig. 7). The lack of ability of Spartina to grow below MSL under mega-tidal conditions suggests that something other than inundation of the sediment is limiting. Mega-tidal conditions expose Spartina to extended periods of leaf inundation under low light that could be lethal. Many marsh plants survive waterlogging by actively moving oxygen to below-ground tissues, whereas inundation of leaves limits gas exchange and therefore can result in rapid and severe rhizome anoxia, particularly at night (Winkel et al. 2011).

This relationship between MTR and the lower limit of marshes has been observed in marsh vegetation more generally. Balke et al. (2016) found that, based on the data set of global marshes, the elevation of the marsh edge relative to MHW is negatively correlated to tidal range with a logarithmic curve. Therefore, they concluded that the potential salt marsh area between the pioneer (lower edge) vegetation elevation and MHW does not proportionally increase with tidal range. The lack of consistency of the lower limit in terms of tidal datums could be due to factors that determine the growth range of salt marsh plants, including flooding and drainage patterns (Mahall and Park 1976; Mendelssohn and Seneca 1980; Armstrong et al. 1985), salinity and freshwater input (Nestler 1977; Niering and Warren 1980; Webb 1983; Zedler 1986; Zedler et al. 1986), soil type and its soil oxidation-reduction (Gray and Bunce 1972; Howes et al. 1981; Mendelssohn et al. 1981), nutrient levels (Valiela and Teal 1974; Osgood and Zieman 1993), physical disturbance (Miller and Egler 1950; Redfield 1972; Bertness and Ellison 1987), and interspecific competition (Pielou and Routledge 1976; Bertness and Ellison 1987). In most studies, information on these factors is limited for comparisons between geographically different marshes. Therefore, it is difficult to predict the growth range of Spartina based on tidal conditions only. Nonetheless, the present study showed that the Spartina marsh distribution of the study sites seems quite close to the predictable elevation of the lower limit in Jindo and Ganghwa tidal condition (Fig. 7). Since 2016 of our study year, most of the Spartina marshes







**Figure 7.** Relationship between mean tidal range and the growth range (**A**), upper limit (**B**) and lower limit (**C**) of occurrence of *Spartina alterniflora* relative to Half Tide Level. The trend line from McKee and Patrick (1988) and 95% Prediction Interval were fitted only to the east coast of United States.

in Jindo and Buno were removed by plowing and mowing, but in Dongmak, *Spartina* patches have been treated primarily on the upper rather than lower tidal flat (KNPS 2016; MOF 2021). Interestingly, the recent satellite imagery (Google Earth 2020, photo provided by Maxar Technologies in 2020) showed that the lowest patches did not spread any more towards the lower elevation than the survey period in 2015. It appears that Dongmak *Spartina* marsh had already reached the lowest possible elevation in 2015. Indeed, it has only been about ten years since



*Spartina* was introduced in Dongmak and Namdong-ri. Therefore, its spread is extremely rapid. However, it should be noted that the *Spartina* vegetation of Buno area had a lag phase for a decade after introduction (unpubl. Data by authors), suggesting that the spread of *Spartina* marsh is not always continuous but also locally different as well.

# Local and regional comparisons

We observed local differences in patch shapes and their vertical distributions, showing many small scattered, round patches in the broadly open tidal flat of Ganghwa area, but rather cohesive and oblong patches in the enclosed bay of Jindo site (Fig. 1, Table 3). Therefore, *Spartina* showed horizontally expanded round patches along the shoreline in Ganghwa tidal flat, but developed atypically, near tidal creeks in a small, enclosed bay in Jindo. Further, Ganghwa site has a gentler slope, a wider and more open tidal flat with a larger tidal range, while Jindo has a relatively steeper slope, a narrower and more enclosed tidal flat with a smaller tidal range. These conditions might also have led to the similar development of patches in two sites of Ganghwa, even though the *Spartina* was introduced at different times, whereas Dongmak and Namdong-ri developed differently even though *Spartina* was introduced almost at the same time in both sites (Table 3, Fig. 4). Consequently, we think that tidal conditions might have determined the vertical distribution of *Spartina* in two study areas, but also topographical differences at the same time may be an important factor to take into consideration in the spatial distribution and form of the *Spartina* patches.

Locality Ganghwa (Dongmak / Buno) Jindo (Namdong-ri) Mean Tidal Range 5.69 m 2.02 m Type of tidal flat opened enclosed bay Size of the study area 4.7 km<sup>2</sup> / 1.1km<sup>2</sup> 0.03 km<sup>2</sup> Slope of the tidal flat gentle steep and undulated Introduction time estimated (unpubl. Data by authors) 2004 / 1990 2004 8,738 m<sup>2</sup> / 2,245 m<sup>2</sup> 6,263 m<sup>2</sup> Invaded area by Spartina marshes as of 2015 atypical, fused Vegetation type many round patches, scattered Main distributional range MHW-MHWN MHWN-MSL Growth range in tidal level (maximum-minimum) below MHWS -below MHWN below AHHW -below MSL

Table 3. Environmental characteristics and vertical distribution in two study localities invaded by Spartina alterniflora in Korea.

# Competition with native salt marsh vegetation

Our results showed that *Spartina* had a potential to grow in any elevation used by native halophytes on the west coast of Korea. This means that the environmental plasticity of *Spartina* associated with immersion time is greater than those of other native halophytes. Competitive vegetation can set the upper limit of *S. alterniflora*, given that it is displaced by congeneric *S. patens* in its native areas of the Eastern U.S. (Bertness 1991). In addition to *S. patens*, common reed marsh *P. australis* is also known to compete with *S. alterniflora* as an invasive species in *Spartina*'s native habitat (Bertness et al. 2002; Vasquez et al. 2006).

An open question is the extent to which *Spartina* will replace or coexist with native marsh vegetation, given that *Spartina* spans its full vertical range (Fig. 4). For instance, non-native *Spartina anglica* along the Wadden Sea coast in Europe was thought to be in a phase of population increase associated with rising water

temperatures (Loebl et al. 2006), with possible ecosystem-level harm. However, it may instead be the case that *S. alterniflora* allows sediment accretion to keep pace with climate-driven sea level rise in a manner promoting local diversity (Granse et al. 2021).

In Ganghwa and Jindo sites, reed marsh and *Spartina* vegetation were found at the same elevation with spatial competition observed in the upper part of the tidal flat (Fig. 5). This phenomenon holds true for the Chinese example of Yangtze River estuary, where *Spartina* has invaded, driving away two native plants, *P. australis* and *S. mariqueter* (Chen et al. 2004; Li et al. 2009).

On the other hand, in the middle part of the tidal flat below MHW in Dongmak, salt marsh has been originally dominated by monospecific S. japonica (Lee et al. 2006), which is a representative low marsh plant in the west coast of Korean peninsula. Bang et al. (2018) studied the distribution of salt marshes in a Siheung tidal flat (37°23'40"N, 126°46'05"E), located about 35 km from the present study site, showing a clear zonation along elevational gradients. Their study area is characterized by deep channels having a thalweg of about 5–6m below the surrounding intertidal surfaces (Wells et al. 1990). The mean tidal range in Siheung tidal flat is 5.57 m, while the vertical distribution of salt marsh plants ranges from 2.4 to 4.2 m above MSL. Four community types occurred in the marsh; (1) Suaeda glauca, Z. sinica, and P. australis at high elevation, (2) Phacelurus latifolius at midhigh elevation, (3) S. japonica at low elevation, (4) Carex scabrifolia at mid-high elevation. At Dongmak in our study, S. japonica occurred primarily from 2.0 to 2.9 m above MSL, whereas its distribution at Siheung was 2.4–4.2 m above MSL with a peak biomass in the elevation between 3.6 m and 3.9 m. S. japonica may be restricted from lower elevations at Siheung tidal flat because the channel flank is unconsolidated fluid-like mud of up to 1 m thickness and dropped sharply (more than 30° slope) to a flat. Suaeda marshes often harbor high macrobenthic biodiversity including two locally abundant crabs Macrophthalmus (Mareotis) japonicus and Cleistostoma dilatatum also found in the bare mud flat in Ganghwa Island (Lee et al. 2016). However, *Suaeda* habitat on the bare mud flat is gradually reduced by the invasion of this exotic plant Spartina (Fig. 5). This effect has also been reported in the Chinese coast, showing that Spartina has restricted the distribution of *Suaeda* spp. and dominated bare mud flats (Wan et al. 2009; Meng et al. 2020). Therefore, we can expect that S. japonica vegetation will be replaced by Spartina marsh, and this competitive displacement could also apply to other surrounding native salt marshes such as T. maritimum and S. triqueter (Figs 4, 5). This displacement can also be predicted for native marsh plants such as S. maritima and Z. sinica in Jindo tidal flat.

# Ecological implications from the introduction of invasive species in Korean coast

The smooth cordgrass *S. alterniflora* with high invasiveness is highly opportunistic and competitive with wide ecological niches. Dispersal pattern and expansion rate of this invasive species appear often inconsistent after the introduction in the recipient waters. However, once their habitat requirements are provided, populations grow rapidly and exponentially.

As of 2020, despite investments in mechanical control of more than US\$3 million over 6 years, *Spartina* continued to expand in Korea. In the US, successful control of widespread (3600 solid ha across 27,000 ha of intertidal area) *Spartina* meadows in Willapa Bay was achieved once an effective herbicide was identified (Major et al. 2003; Knott et al. 2013; Patten et al. 2017).



Worldwide, *Spartina* has become a dominant part of tidal flats where it has invaded, which is likely to occur on Korean tidal flats around Ganghwa and Jindo Islands. Recently, in Chinese coasts, the native intertidal halophyte *Suaeda* spp. and even dwarf eelgrass *Zostera japonica* have been degraded significantly by *Spartina* invasion in the Yellow River Delta. *Spartina* marsh competes with native plants, threatens native ecosystems and coastal aquaculture, and causes local biodiversity to decline (He et al. 2007; Wan et al. 2009; Wan et al. 2014; Liu et al. 2018; Meng et al. 2020; Yue et al. 2021).

We also vividly witnessed this competitive effect and rapid expansion of the *Spartina* invasion on native halophytes in Korean intertidal wetlands (unpubl. Data by authors). In addition, this invasive *Spartina* greatly alters the rates and pathways of organic carbon oxidation and associated microbial communities (An et al. 2020), but also increases belowground biomass and decreases macrofaunal density and diversity in the same study area of Ganghwa intertidal Wetland located at the Han River Estuary, Korea (Shin et al. 2022).

As *Spartina* increased at Dongmak from 2010 to 2015, *Suaeda japonica* marsh decreased its area from 99,229 m<sup>2</sup> to 64,986 m<sup>2</sup>. This locally important and dominant native halophyte harbors the highest diversity in macrofauna among habitat types on these tidal flats (Lee et al. 2016; Shin et al. 2022). In addition, bare mud-flat-based traditional fishing activities associated with previous local habitats will also be excluded by *Spartina* invasion.

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# Authors' contributions

Sungtae Kim: Research Conceptualization, Sample design and Methodology, Investigation and Data collection, Data analysis and Interpretation, Resources, Writing- Original Draft, Visualization.

Jae-Sang Hong: Research Conceptualization, Sample design and Methodology, Investigation and Data collection, Data analysis and Interpretation, Resources, Writing- Original Draft and Review, Visualization, Funding provision

Cheol Yu: Investigation and Data collection, Writing, Resources.

Jennifer Ruesink: Research Conceptualization, Sample design and Methodology, Writing- Review & Editing, Visualization.

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#### References

- An SQ, Gu BH, Zhou CF, Wang ZS, Deng ZF, Zhi YB, Li HL, Chen L, Yu DH, Liu YH (2007) Spartina invasion in China: Implications for invasive species management and future research. Weed Research 47(3): 183–191. https://doi.org/10.1111/j.1365-3180.2007.00559.x
- An S-U, Cho H, Jung U-J, Kim B, Lee H, Hyun J-H (2020) Invasive *Spartina anglica* greatly alters the rates and pathways of organic carbon oxidation and associated microbial communities in an intertidal wetland of the Han River Estuary, Yellow Sea. Frontiers in Marine Science 7: 59. https://doi.org/10.3389/fmars.2020.00059
- Armstrong W, Wright EJ, Lythe S, Gaynard TJ (1985) Plant zonation and the effects of the springneap tidal cycle on soil aeration in a humber salt marsh. The Journal of Ecology 73(1): 323–339. https://doi.org/10.2307/2259786
- Balke T, Stock M, Jensen K, Bouma TJ, Kleyer M (2016) A global analysis of the seaward salt marsh extent: The importance of tidal range. Water Resources Research 52(5): 3775–3786. https://doi.org/10.1002/2015WR018318
- Bang JH, Bae M-J, Lee EJ (2018) Plant distribution along an elevational gradient in a macrotidal salt marsh on the west coast of Korea. Aquatic Botany 147: 52–60. https://doi.org/10.1016/j. aquabot.2018.03.005
- Bertness MD (1991) Zonation of *Spartina patens* and *Spartina alterniflora* in a New England salt marsh. Ecology 72(1): 138–148. https://doi.org/10.2307/1938909
- Bertness MD, Ellison AM (1987) Determinants of pattern in a New England salt marsh plant community. Ecological Monographs 57(2): 129–147. https://doi.org/10.2307/1942621
- Bertness MD, Ewanchuk PJ, Silliman BR (2002) Anthropogenic modification of New England salt marsh landscapes. Proceedings of the National Academy of Sciences 99(3): 1395–1398. https://doi.org/10.1073/pnas.022447299
- Blum MJ, Jun Bando K, Katz M, Strong DR (2007) Geographic structure, genetic diversity and source tracking of *Spartina alterniflora*. Journal of Biogeography 34(12): 2055–2069. https://doi. org/10.1111/j.1365-2699.2007.01764.x
- Bruno JF, Kennedy CW (2000) Patch-size dependent habitat modification and facilitation on New England cobble beaches by *Spartina alterniflora*. Oecologia 122: 98–108. https://doi. org/10.1007/PL00008841
- Byers SE, Chmura GL (2007) Salt marsh vegetation recovery on the bay of fundy. Estuaries and Coasts 30(5): 869–877. https://doi.org/10.1007/BF02841340
- Callaway JC, Josselyn MN (1992) The introduction and spread of smooth cordgrass (*Spartina alterni-flora*) in South San Francisco Bay. Estuaries 15(2): 218–226. https://doi.org/10.2307/1352695
- Chapman VJ (1974) Salt marshes and salt deserts of the world. In: Reimold RJ, Queen WH (Eds) Ecology of Halophytes. Academic Press, New York, 3–19. https://doi.org/10.1016/b978-0-12-586450-3.50006-8
- Chen Z, Li B, Zhong Y, Chen J (2004) Local competitive effects of introduced *Spartina alterniflora* on *Scirpus mariqueter* at Dongtan of Chongming Island, the Yangtze River estuary and their potential ecological consequences. Hydrobiologia 528: 99–106. https://doi.org/10.1007/s10750-004-1888-9
- Google Earth (2020) Google Earth. https://earth.google.com/web/@37.58970752,126.44697 401,783.27507602a,0d,35y,-0.4764h,23.7291t,359.9986r?utm\_source=earth7&utm\_campaign=vine&hl=ko [Accessed 14 December 2022]
- Granse D, Suchrow S, Jensen K (2021) Long-term invasion dynamics of *Spartina* increase vegetation diversity and geomorphological resistance of salt marshes against sea level rise. Biological Invasions 23: 871–883. https://doi.org/10.1007/s10530-020-02408-0
- Gray AJ (1992) Saltmarsh plant ecology: zonation and succession revisited. In: Allen JRL, Pye K (Eds) Saltmarshes: Morphodynamics, Conservation and Engineering Significance. Cambridge University Press, Cambridge, 63–79.



- Gray AJ, Bunce RGH (1972) The Ecology of Morecambe Bay. VI. Soils and Vegetation of the Salt Marshes: A Multivariate Approach. The Journal of Applied Ecology 9(1): 221–234. https://doi. org/10.2307/2402058
- He W, Feagin R, Lu J, Liu W, Yan Q, Xie Z (2007) Impacts of introduced *Spartina alterniflora* along an elevation gradient at the Jiuduansha Shoals in the Yangtze Estuary, suburban Shanghai, China. Ecological Engineering 29(3): 245–248. https://doi.org/10.1016/j.ecoleng.2006.08.004
- Howes BL, Howarth RW, Teal JM, Valiela I (1981) Oxidation-reduction potentials in a salt marsh: Spatial patterns and interactions with primary production. Limnology and Oceanography 26(2): 350–360. https://doi.org/10.4319/lo.1981.26.2.0350
- Jung S-Y, Park S-H, Lee K-H, Yang J-C, Chang K-S, Chung J-M, Choi K, Shin C-H, Lee Y-M (2015) A potential risk of invasive alien plants of gen. *Spartina* (Poaceae) in South Korea. Proceedings 46<sup>th</sup> Annual Meeting Korean Society of Plant Taxonomists. Incheon, Korea, February 5, 2015. Korean Society of Plant Taxonomists, Pocheon, Republic of Korea, 49 pp. [in Korean]
- KHOA (2017) Annual report of Korea oceanographic observation network 2016. Korea Hydrographic and Oceanographic Agency, Republic of Korea, 342 pp. [in Korean]
- Kim E-K, Kil J, Joo Y-K, Jung Y-S (2015) Distribution and botanical characteristics of unrecorded alien weed *Spartina anglica* in Korea. Weed & Turfgrass Science 4(1): 65–70. [in Korean] https://doi.org/10.5660/WTS.2015.4.1.65
- Kimura T, Hanai T, Kimura S, Fujioka E (2016) Identification of invasive alien species *Spartina alterniflora* in Japan using morphological characteristics as compared with native species *Phragmites australis*. Japanese Journal of Benthology 70(2): 91–94. [in Japanese] https://doi.org/10.5179/ benthos.70.91
- Knott CA, Webster EP, Nabukalu P (2013) Control of smooth cordgrass (*Spartina alterniflora*) seedlings with four herbicides. Journal of Aquatic Plant Management 51: 132–135.
- KNPS (2016) Eradication of the ecosystem disturbing species from twenty National Parks. Korea National Park Service, July 11, 2016. [in Korean] https://www.knps.or.kr/front/portal/open/ pnewsDtl.do?menuNo=8000319&pnewsId=PNEWSM007123 [Accessed 14 December 2022]
- Lee H-G, Park H-S, Hong J-S, Je J-G, Lee J-H (2006) Spatio-temporal variation in the benthic environmental conditions and salt marsh vegetation in Donggeomdo, Incheon, Korea. Korean Journal of Fisheries and Aquatic Sciences 39: 180–188. https://doi.org/10.5657/kfas.2006.39.spc1.180
- Lee H-G, Yoon K-T, Park H-S, Hong J-S, Lee J-H (2016) The influence of environmental variables on distribution of macrobenthic community in salt marsh vegetation in Donggeomdo, Ganghwa on the west coast of Korea. Ocean and Polar Research 38(2): 115–128. [in Korean] https://doi. org/10.4217/OPR.2016.38.2.115
- Lee Y-K, Ryu J-H, Choi J-K, Soh J-G, Eom J-A, Won J-S (2011) A study of decadal sedimentation trend changes by waterline comparisons within the Ganghwa tidal flats initiated by human activities. Journal of Coastal Research 27(5): 857–869. https://doi.org/10.2112/JCOAS-TRES-D-10-00150.1
- Li B, Liao CH, Zhang XD, Chen HL, Wang Q, Chen ZY, Gan XJ, Wu JH, Zhao B, Ma ZJ, Cheng XL, Jiang LF, Chen JK (2009) *Spartina alterniflora* invasions in the Yangtze River estuary, China: An overview of current status and ecosystem effects. Ecological Engineering 35(4): 511–520. https://doi.org/10.1016/j.ecoleng.2008.05.013
- Liu M, Mao D, Wang Z, Li L, Man W, Jia M, Ren C, Zhang Y (2018) Rapid invasion of *Spartina alterniflora* in the coastal zone of mainland China: New observations from Landsat OLI images. Remote Sensing 10(12): 1933. https://doi.org/10.3390/rs10121933
- Liu W, Zhang Y, Chen X, Maung-Douglass K, Strong DR, Pennings SC (2020) Contrasting plant adaptation strategies to latitude in the native and invasive range of *Spartina alterniflora*. New Phytologist 226: 623–634. https://doi.org/10.1111/nph.16371
- Loebl M, van Beusekom JEE, Reise K (2006) Is the spread of the neophyte *Spartina anglica* recently enhanced by increasing temperatures? Aquatic Ecology 40: 315–324. https://doi.org/10.1007/s10452-006-9029-3



- Maebara Y, Tamaoki M, Iguchi Y, Nakahama N, Hanai T, Nishino A, Hayasaka D (2020) Genetic diversity of invasive *Spartina alterniflora* Loisel. (Poaceae) introduced unintentionally into Japan and its invasion pathway. Frontiers in Plant Science 11: 556039. https://doi.org/10.3389/ fpls.2020.556039
- Mahall BE, Park RB (1976) The Ecotone Between *Spartina foliosa* Trin. and *Salicornia virginica* L. in Salt Marshes of Northern San Francisco Bay: III. Soil Aeration and Tidal Immersion. The Journal of Ecology 64(3): 811–819. https://doi.org/10.2307/2258810
- Major WW, Grue CE, Grassley JM, Conquest LL (2003) Mechanical and chemical control of smooth cordgrass in Willapa Bay, Washington. Journal of Aquatic Plant Management 41: 6–12.
- McKee KL, Patrick WH (1988) The relationship of smooth cordgrass (*Spartina alterniflora*) to tidal datums: A review. Estuaries 11: 143–151. https://doi.org/10.2307/1351966
- Mendelssohn IA, Seneca ED (1980) The influence of soil drainage on the growth of salt marsh cordgrass Spartina alterniflora in North Carolina. Estuarine and Coastal Marine Science 11(1): 27–40. https://doi.org/10.1016/S0302-3524(80)80027-2
- Mendelssohn IA, McKee KL, Patrick WH (1981) Oxygen deficiency in *Spartina alterniflora* roots: Metabolic adaptation to anoxia. Science 214(4519): 439–441. https://doi.org/10.1126/science.214.4519.439
- Meng W, Feagin RA, Innocenti RA, Hu B, He M, Li H (2020) Invasion and ecological effects of exotic smooth cordgrass *Spartina alterniflora* in China. Ecological Engineering 143(15): 105670. https://doi.org/10.1016/j.ecoleng.2019.105670
- Miller WR, Egler FE (1950) Vegetation of the Wequetequock-Pawcatuck Tidal-Marshes, Connecticut. Ecological Monographs 20(2): 143–172. https://doi.org/10.2307/1943548
- MOF (2021) National Investigation of Marine Ecosystem 2021 Marine Ecology Series The West Sea and Western Part of the South Sea. Ministry of Oceans and Fisheries, Republic of Korea, 342 pp. [in Korean]
- Nestler J (1977) Interstitial salinity as a cause of ecophenic variation in *Spartina alterniflora*. Estuarine and Coastal Marine Science 5(6): 707–714. https://doi.org/10.1016/0302-3524(77)90043-3
- Niering WA, Warren RS (1980) Vegetation patterns and processes in new england salt marshes. BioScience 30(5): 301–307. https://doi.org/10.2307/1307853
- Osgood DT, Zieman JC (1993) Factors controlling aboveground *Spartina alterniflora* (Smooth cordgrass) tissue element composition and production in different-age barrier island marshes. Estuaries 16(4): 815–826. https://doi.org/10.2307/1352440
- Patten K, O'Casey C, Metzger C (2017) Large-scale chemical control of smooth cordgrass (*Spartina alterniflora*) in Willapa Bay, WA: Towards eradication and ecological restoration. Invasive Plant Science and Management 10(3): 284–292. https://doi.org/10.1017/inp.2017.25
- Pennings SC, Bertness MD (2001) Salt marsh communities. In: Bertness MD, Gaines SD, Hay ME (Eds) Marine community Ecology. Sinauer, Massachusetts, 289–316.
- Pielou EC, Routledge RD (1976) Salt marsh vegetation: Latitudinal gradients in the zonation patterns. Oecologia 24: 311–321. https://doi.org/10.1007/BF00381137
- Redfield AC (1972) Development of a New England Salt Marsh. Ecological Monographs 42(2): 201–237. https://doi.org/10.2307/1942263
- Ruesink J (2018) Biological Invasions of Mudflats. In: Beninger P (Eds) Mudflat Ecology. Aquatic Ecology Series, Volume 7. Springer, Cham, 271–308. https://doi.org/10.1007/978-3-319-99194-8\_11
- Sayce K (1988) Introduced cordgrass, Spartina alterniflora Loisel., in salt marshes and tidelands of Willapa Bay, Washington. US Fish and Wildlife Service, Willapa National Wildlife Refuge. US Fish and Wildlife Service contract No FWSI87058(TS).
- Shin W, Oh M, Hong J-S, Byun C, Lee EJ (2022) Early invasion of common cordgrass (*Spartina anglica*) increases belowground biomass and decreases macrofaunal density and diversity in a tidal flat marsh. Biological Invasions 24: 3615–3629. https://doi.org/10.1007/s10530-022-02866-8
- Simenstad C, Thom R (1995) *Spartina alterniflora* (smooth cordgrass) as an invasive halophyte in Pacific northwest estuaries. Hortus Northwest 6(1): 9–40.



- Strong DR, Ayres DR (2013) Ecological and evolutionary misadventures of *Spartina*. Annual review of ecology, evolution, and systematics 44(1): 389–410. http://doi.org/10.1146/annurev-ecol-sys-110512-135803
- Valiela I, Teal JM (1974) Nutrient limitation in salt marsh vegetation. In: Reimold RJ, Queen WH (Eds) Ecology of Halophytes. Academic Press, New York, 547–563. https://doi.org/10.1016/ B978-0-12-586450-3.50025-1
- Vasquez EA, Glenn EP, Guntenspergen GR, Brown JJ, Nelson SG (2006) Salt tolerance and osmotic adjustment of *Spartina alterniflora* (Poaceae) and the invasive M haplotype of *Phragmites australis* (Poaceae) along a salinity gradient. American Journal of Botany 93(12): 1784–1790. http://doi.org/10.3732/ajb.93.12.1784
- Wan H, Wang Q, Jiang D, Fu J, Yang Y, Liu X (2014) Monitoring the invasion of *Spartina alterni-flora* using very high resolution unmanned aerial vehicle imagery in Beihai, Guangxi (China). The Scientific World Journal 2014: 638296. http://doi.org/10.1155/2014/638296
- Wan S, Qin P, Liu J, Zhou H (2009) The positive and negative effects of exotic Spartina alterniflora in China. Ecological Engineering 35(4): 444–452. http://doi.org/10.1016/j.ecoleng.2008.05.020
- Webb JW (1983) Soil water salinity variations and their effects on *Spartina alterniflora*. Contributions in marine science 26: 1–14. http://hdl.handle.net/2152/18033
- Wells JT, Adams Jr CE, Park YA, Frankenberg EW (1990) Morphology, sedimentology and tidal channel processes on a high-tide-range mudflat, west coast of South Korea. Marine Geology 95(2): 111–130. https://doi.org/10.1016/0025-3227(90)90044-K
- Winkel A, Colmer TD, Pedersen O (2011) Leaf gas films of *Spartina anglica* enhance rhizome and root oxygen during tidal submergence. Plant, Cell & Environment 34: 2083–2092. https://doi.org/10.1111/j.1365-3040.2011.02405.x
- Yue S, Zhou Y, Xu Shaochun, Zhang X, Liu M, Qiao Y, Gu R, Xu Shuai, Zhang Y (2021) Can the non-native salt marsh halophyte *Spartina alterniflora* threaten native seagrass (*Zostera japonica*) habitats? A case study in the Yellow River Delta, China. Frontiers in Plant Science 12: 643425. http://doi.org/10.3389/fpls.2021.643425
- Zedler JB (1986) Catastrophic flooding and distributional patterns of Pacific cordgrass (*Spartina foliosa* Trin.). Bulletin of the Southern California Academy of Sciences 85(2): 74–86.
- Zedler JB, Covin J, Nordby C, Willia P, Boland J (1986) Catastrophic events reveal the dynamic nature of salt-marsh vegetation in Southern California. Estuaries 9(1): 75–80. https://doi. org/10.2307/1352195
- Zhang RS, Shen YM, Lu LY, Yan SG, Wang YH, Li JL, Zhang ZL (2004) Formation of Spartina alterniflora salt marshes on the coast of Jiangsu Province, China. Ecological Engineering 23(2): 95–105. https://doi.org/10.1016/j.ecoleng.2004.07.007